# Acoustic-based estimation of fish stocks in Widas Reservoir, East Java, Indonesia 

Siti Nurul Aida ${ }^{1}$, Agus Djoko Utomo ${ }^{1}$, Safran Makmur ${ }^{1}$, Tuah Nanda M. Wulandari ${ }^{1}$, Khoirul Fatah ${ }^{1}$, Yosmaniar ${ }^{2}$, Indra Suharman ${ }^{3}$, Ulung Jantama Wisha ${ }^{4,{ }^{*}}$<br>${ }^{1}$ Research Center for Conservation of Marine and Inland Water Resources, National Research and Innovation Agency (BRIN), Cibinong 16911, Indonesia<br>${ }^{2}$ Research Center for Marine and Land Bioindustry, National Research and Innovation Agency (BRIN), Cibinong 16911, Indonesia<br>${ }^{3}$ Department of Aquaculture, Faculty of Fisheries and Marine Sciences, Universitas Riau, Pekanbaru-Riau 28292, Indonesia<br>${ }^{4}$ Research Center for Oceanography, National Research and Innovation Agency (BRIN), Jakarta 14430, Indonesia


#### Abstract

Widas Reservoir is situated in an area of 570 ha in the Pajaran Village, Madiun Regency, East Java Province, Indonesia, playing an essential role in fisheries, with the average fish catch per year of about 283 tons/year. This study explores the standing stock, growth parameters, mortality, and exploitation rates of several dominant fishes in Widas Reservoir. This study was carried out from February to November 2019. Fish stocks were estimated using acoustic tools, fish catch records, and sizes collected by local enumerators. Fish length frequency sampling was conducted on several dominant fish species, such as Oreochromis niloticus, Barbonymus gonionotus, and Osteochilus vittatus. Based on the length-frequency data, estimating fish population dynamics, the fish population dynamics (infinitive length $\left(L_{\infty}\right)$ and growth coefficient $(K)$ ) estimation was run in a time series using the Fish Sock Analysis Tool, II (FISAT II) program package. Moreover, the estimation of natural mortality parameters, the fishing mortality parameter, and the exploitation rate was also performed. The approximated overall fish stock in the Widas Reservoir was about $79,848 \mathrm{~kg}$, which lowered with the increase in water depth. Of particular concern, in the surface layer at a depth between $1-5 \mathrm{~m}$, the fish stock reached $58,813 \mathrm{~kg}$, while in the deeper zone (> 15 m ), the value significantly lowered by about $98 \%$, reaching 1,219 kg . These results indicate an overfishing in the Widas Reservoir. The value of the exploitation rate (E) of B. gonionotus was 0.748 , $O$. niloticus 0.8 , and $O$. vittatus 0.7 , respectively, proving the overfishing states occurred in the study area. Therefore, regulations governing the number of catches and the use of fishing gear are crucial in Widas Reservoir, particularly the use of lift and gill nets with a mesh size of less than 2 cm .


Keywords: Stock assessment, Fish stock, Mortality, Overfishing, Widas Reservoir

[^0]
## Introduction

In Indonesia, there are about 102 large and small reservoirs, $80 \%$ of which are on Java (Hart et al., 2002). The number of large reservoirs ( $>500 \mathrm{ha}$ ) is around $15 \%$, and the rest ( $85 \%$ ) are small reservoirs. East Java has about 21 reservoirs consisting of two large reservoirs, namely Karangkates and Wonorejo Reservoirs, and 19 small reservoirs, including Widas Reservoir. Widas Reservoir covers an area of 500 ha, is a multipurpose reservoir with the primary role in irrigating rice fields and produces about 650 kW of electric power (Prabowo \& Nawiyanto, 2020). Other parts are drinking water sources, tourism, aquaculture, and capture fisheries. This reservoir has the potential to increase fishery production.

The damming of the Widas River, the Berantas Sub-Watershed, and the water source from Mount Wilis arranged the Widas Reservoir. Aside from being subjected to fishing pressure, inland waters are also threatened by local pollution, for instance, microplastic pollution (MP) in the Brantas River (Buwono et al., 2021). However, this issue does not directly damage Widas Reservoir waters, even though the inlet is the Bening River (Bening as transparent), a tributary of the Brantas River. Widas Reservoir provides drinking water (Aida et al., 2018). Another name for Widas Reservoir is Bening Dam, which is 40 km from Madiun City. The average catch is about 283 tons/ year, consisting of the following fish species: Cyprinus carpio, Barbonymus gonionotus, Oreochromis niloticus, Pangasianodon hypophthalmus, Notopterus notopterus, Osteochilus vittatus, and Rasbora spp. (Aida et al., 2018).

Fishing activities in Widas Reservoir are very intensive. Several main fishing gears are used daily: cast net, gillnet, pot traps, hook line, scop net, and lift net. Gillnets have recently skyrocketed, mainly operated in reservoir inlets at the beginning of the rainy season to catch spawned fish. The average fish caught in Widas Reservoir yielded from the use of gillnets are approximately $5 \mathrm{~kg} /$ day/person, cast net about $3 \mathrm{~kg} /$ day/day, pot traps $3 \mathrm{~kg} /$ day/person, hook line $2 \mathrm{~kg} /$ day/person, and scop net of about $1 \mathrm{~kg} /$ day $/$ person. The catch through fishing is insignificant in the Widas Reservoir because of limited fish resources and the increasing number of fishers. Several fish species are rarely found in Widas Reservoir, such as Pangasius, Notopterus, and Clarias (Aida et al., 2018). The main problem is overexploitation, whereby the immature fish, particularly immaturity in the gonad, commence to be caught by locals. For instance, Minnows fish (B. gonionotus), with a size of 16 cm , have been
caught intensively, despite the gonad maturity being ready when the size is 28 cm . Therefore, the reproduction cycle is hampered. A similar state is also suffered by Cichlids ( $O$. niloticus) and Barbs fish (O. vittatus) (Aida et al., 2018, 2022a).

Fish stocks in the waters can be lowered if overfishing occurs, decreasing fish populations and even extinction. Fish stock assessment research aims to provide a threshold for optimal utilization of fish resources. Fishing activities should be reasonable, whereby the number of catches should equal the amount of fish recruitment (Sparre \& Venema, 1992). Many fishing gear are being operated inland in Indonesian waters, including small and large gears (hook line, cast net, pot traps, barrier trap, filtering device, active seine)-the considerable increase in human population results in the elevated number of local fishers. Therefore, the pressure on the sustainability of fish resources and overexploitation undergoes an upward trend (Utomo \& Samuel, 2017).

Widas Reservoir is essential for supporting the fish stock and can contribute to the local community's economy. Concerning fish sustainability, the number of fish caught in the Widas Reservoir needs to be further controlled. A study of fish stocks is crucial in managing fish resources. Estimating fish stock using hydroacoustic methods has been applied in all aquatic ecosystems (Tsagarakis et al., 2015). Hydroacoustic surveys are increasingly being used to assess the distribution and abundance of freshwater fish. One of the drawbacks of this method is that fish can avoid moving boats, potentially causing bias in the assessment of fish stocks (Wheeland \& Rose, 2014). The application of hydroacoustic has grown after being established as a standard method for fish stock assessment (Perivolioti et al., 2020). To date, previous studies in the East Java Reservoir only concern aquatic environmental problems, and a few studies assessed the fish stocks. For instance, Aida \& Utomo (2018) examined the carrying capacity for supporting floating net cages in Pondok Reservoir, East Java. On the other hand, the arrangement of sluice gates in reservoirs for rice fields has been studied by Susilo (2019). Viani \& Retnaningdyah (2018) surveyed the variability of water quality in East Java's Lahor Reservoir. Relating to the fish stock, Widas Reservoir is stocked annually (Utomo et al., 2019). However, no further assessments based on the stocked fish data have been analyzed, particularly the sustainability assessment and the overexploitation status. Thus, these aspects should be investigated to provide the recent status and input for the sustainable management of fish resources. This study aims to unravel the standing stock, growth pa-
rameters, mortality, and exploitation rates of several dominant fish in Widas Reservoir.

## Materials and Methods

## Research location and time

The research was carried out at the Widas Reservoir in East Java, Indonesia, from February to November 2019. The research consisted of a field survey, stock estimation experiments based on acoustic data, and collecting fish catch and size data from enumerators. The fish catch data was collected every day, while the measurement of fish size was conducted every two weeks. The survey was conducted thrice in March, June, and September 2019 to collect data from enumerators and fish sampling. Meanwhile, a survey of fish stock assessment using acoustic instruments was conducted in June 2019. Based on preliminary information about fishing activities, it is suspected that the status of fish stocks in the Widas Reservoir has been overfishing.

## Estimation of fish growth and mortality

It is necessary to input data on fish length frequency in a time series to estimate growth and mortality parameters. The length frequency distribution was carried out on several dominant fish species in the Widas Reservoir, such as O. niloticus, B. gonionotus, and O. vittatus. Fish samples were obtained from various fishing gears, with the help of local fishers as enumerators. The fish caught were taken randomly with a scoop net, then the total length of the fish was measured (cm). The recording of fish catch data was performed every day, while the measurement of fish length was carried out every two weeks.

Growth parameters were estimated using the Von Bertalanffy growth equation as follows:

$$
\begin{equation*}
L_{t}=L_{\infty}\left(e^{-k\left(t-t_{0}\right)}\right) \tag{1}
\end{equation*}
$$

where:
$L_{t}=$ the length at age
$L_{\infty}=$ the asymptotic length
$K=$ the growth factors
$t_{0}=$ the theoretical age when the size of fish is zero.
Data analysis used the Electric Length Frequency Analysis (ELEFAN) program (FAO, Rome, Italy) in the Fish Sock Analysis Tool, II (FISAT II) program package (FAO, Rome, Italy;

Gayanilo et al., 1995). The total mortality parameter (Z) was analyzed using the Jones and Van Zalinge method (Sparre \& Venema, 1992), which was based on group length and growth parameters obtained as follows:

$$
\begin{equation*}
\log C\left\{\left(L, L_{\infty}\right)\right\}=a+\frac{Z\left(\log \left(L_{\infty}-L\right)\right.}{K} \tag{2}
\end{equation*}
$$

where:

$$
\begin{aligned}
& C\left(L, L_{\infty}\right)=\text { Cumulative catch at length } L(\mathrm{~cm}) \\
& L_{\infty}=\text { Infinity length } \\
& K=\text { Growth constant } \\
& \mathrm{Z}=\text { Total mortality }
\end{aligned}
$$

On the other hand, natural mortality parameters (M) based on Pauly's empirical equation was also calculated (Pauly, 1984):

$$
\begin{align*}
\log (\mathrm{M})= & -0.0152-0.2790 \log \left(L_{\infty}\right)+0.6543 \log (K) \\
& +0.4634 \log (T) \tag{3}
\end{align*}
$$

Fishing mortality parameters $(F)=Z-M$, and fishing rate $E=F / Z,(E>0.5$ indicates overfishing $)($ Pauly, 1984).

## Acoustic-based estimation of fish stocks and bathymetry

Bathymetry and fish stock surveys were conducted in June 2019 using an acoustic device SIMRAD EY-60 (SIMRAD, Notteroy, Norway; portable echosounder) with a 120 kHz transducer and a global positioning system (GPS) tool. The transducer was installed and mounted on the right side of the boat with a 3 gross tone (GT) engine. The position of the transducer was vertical, pointing down so that the vertical beams could provide a representative approximation for the water column. The echosounder was connected to a portable computer, which offers real-time viewing and data storage on the hard disk.

Scientific echosounders can estimate the intensity of the echo/reflection of sound from the fish's body to the echosounder (target strength, TS). Several models describing the relationship between fish size and TS should be used to detect fish using Single Echo Detection (SED) as a source of average target power (TS) to evaluate fish density (Lian et al., 2017). The second step is converting fish size into biomass (weight) by applying the length-weight relationship. The echo data was processed with Sonar 5-Pro and analyzed using Microsoft Excel 2019.

Real-time data acquisition was performed using the ER60 software. Further acoustic data processing was conducted using

SONAR-4 software (Cakewalk, Boston, MA, USA; Table 1). The survey pattern followed zigzag transects in the reservoir waters representing the types of reservoir water bodies such as bays, middle of reservoirs, and reservoir edges (Fig. 1). We also sampled fish composition from multi-fishing gear, including gillnets from various mesh sizes, cast nets, and lift nets.

Estimating fish species composition was performed by recording fish catches by several fishers as enumerators. The fishing gear used is nonselective fishing gear that can catch fish
of various species and sizes. The length frequency data of fish was approached by measuring the total length of fish every two weeks, whereas the fish catch data were recorded daily by the enumerator using a specific form. The catch data recorded includes the type of fishing gear used, the time of catching, the amount of catch (kg), and the composition of fish species (\%).

All data retrieved from the echo sounder should be stored on the computer's hard disk for further processing using Simrad ER60 software (Renfree \& Demer, 2016). The acoustic data were

Table 1. Acoustic equipment for fish stock estimation

| No Equipment type | Function |  |
| :--- | :--- | :--- |
| 1 | Acoustics: Portable Scientific Echosounder SIMRAD EY-60, 120 kHz transducer | The main tool for estimating fish stocks |
| 2 | One Gienset Portable Unit, 1,000 Watt | Power source during the research |
| 3 | Wooden boat, 3 GT engine | Equipment carrier, installation site acoustic instrument beside the boat body |
| 4 | Laptop, memory $>2 \mathrm{~GB}$, hard drive $>80 \mathrm{~GB}$ | Data acquisition during operation in the boat. |
| 5 | Personal computer (PC), memory $>2 \mathrm{~GB}$, hard drive $>80 \mathrm{~GB}$ | Post-processing data |
| 6 | ER60 software | Processing echogram data into data-threshold (data acquisition). |
| 7 | SONAR-4 software | Advanced data processing, stored in ASCII |
| 8 | Microsoft excel software | Tabulating data |

[^1]

Fig. 1. The track of hydroacoustic survey in Widas Reservoir, East Java.
processed using ECHOVIEW version 5, yielding a backscattering coefficient (coefficient of fish cross-section) and distribution of single fish TS values in decibels ( dB ) as a reflection index of fish size (Mello \& Rose, 2009). TS is the strength of a target (fish) to reflect acoustic sounds and has a close relationship with the size of the fish. The larger the size of the fish, the greater the TS value. After gaining the values and characteristics of the fish's TS, fish stocks and abundant resources can be estimated. According to Hannachi et al. (2004), the relationship between TS and óbs (backscattering cross-section, $\mathrm{mm}^{2}$ ) can be calculated by converting the TS value to length (L) for fish using the following equation:

$$
\begin{equation*}
T S=20 \log L-73.97 \tag{4}
\end{equation*}
$$

For converting fish size into fish weight using the equation as follows:

$$
\begin{equation*}
W=a L^{b} \tag{5}
\end{equation*}
$$

where:
$\mathrm{L}=$ Total length (cm)
$\mathrm{W}=$ Total weight $(\mathrm{g})$

Then all existing individual fish sizes can be changed into fish biomass.

Maximum Sustainable Yield (MSY) $=0.5 \times Z \times$ Biomass (6)
Total Allowable Catch $(T A C)=80 \% \times M S Y$

## Results

## Fish population parameters

An assessment of the population dynamics of the dominant fish often caught has been performed to determine the status of fish stocks in the Widas Reservoir. The dominant fish are B. gonionotus, O. niloticus and $O$. vittatus. Based on the length frequency data analysis of the dominant species, the maximum length $\left(L_{\infty}\right)$, growth $(K)$, natural mortality $(M)$, total mortality $(Z)$, fishing mortality $(F)$, and exploitation rate $(E)$ has been calculated as presented in Table 2.

## Tawes/Minnows (Barbonymus goniotus)

The length distribution (cm) of Tawes/Minnows fish was between $8.0-36 \mathrm{~cm}$, and the number of samples ( $n$ ) taken for the analysis was 860 fish (Fig. 2). Based on the analysis of fish size, the infinitive length $\left(L_{\infty}\right)=45 \mathrm{~cm}, K=0.63$ year $^{-1}$ and $t_{0}=-0.23$ year, the equation for the growth of Vont Bertalanffy: $L_{t}=45$ (1 $\left.-e^{-0.63(t+0.23)}\right)$ (Fig. 3). Assuming the average water temperature is $29.5^{\circ} \mathrm{C}$, based on an empirical equation, natural mortality $(M)=$ 1.24 year $^{-1}$. Based on Jones and Van Zalinge Plot analysis on the

Table 2. Some population parameters of Bonionotus gonionotus, Osteochilus vittatus, and Oreochromis niloticus in the Widas Reservoir Madiun Regency, East Java

| Fish species | $L_{\infty}(\mathrm{cm})$ | $K=\left(\right.$ year $\left.^{-1}\right)$ | $M=\left(\right.$ year $\left.^{-1}\right)$ | $Z=\left(\right.$ year $\left.^{-1}\right)$ | $F=\left(\right.$ year $\left.^{-1}\right)$ | $E$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| B. gonionotus | 45 | 0.63 | 1.24 | 4.78 | 3.54 | 0.74 |
| O. niloticus | 44.4 | 0.7 | 1.30 | 7.55 | 6.25 | 0.83 |
| O. vittatus | 29 | 0.55 | 1.25 | 3.70 | 2.45 | 0.66 |



Fig. 2. The length frequency data of Tawes/Minnows (Bonionotus gonionotus) in the Widas Reservoir, East Java.


Fig. 3. Growth of Tawes/Minnows (Bonionotus gonionotus) in Widas Reservoir, East Java.
size composition of Tawes/Minnows, the results obtained were total mortality $(Z)=4.78$ year $^{-1}$, the fishing mortality parameter was $(F)=3.57$ year $^{-1}$ and the exploitation rate $(E)=0.74$. The fishing mortality $(F)$ is greater than the natural mortality $(M)$, and the exploitation rate $E>0.5$ is an indicator of overfishing.

## Nila/Cichlid (Oreochromis niloticus)

The size of the Nila/Cichlid sample ranged from 11.1 to 35 cm . The number of samples ( $n$ ) was 4,452 fish (Fig. 4). Based on the analysis of the length-frequency data, the infinitive length was $\left(L_{\infty}\right)=44.4 \mathrm{~cm}, K=0.7$ year $^{-1}$ and $t_{0}=-0.21$ year and the growth equation $L_{t}=44.4\left(1-e^{-0.7(t+0.21)}\right)$ (Fig. 5). Assuming an average temperature of $29.5 \pm 5^{\circ} \mathrm{C}$, based on the empirical equation, the natural mortality parameter value was $(M)=1.30$ year $^{-1}$. Based on the Jones and Van Zalinge plot analysis on the size composition of Nila/Cichlid fish during the study, the results obtained were total mortality $(Z)=7.55$ year $^{-1}$, fishing mortality $(F)=6.25$ year ${ }^{-1}$, and the exploitation rate $(E)=0.83$. The fishing mortality $(\mathrm{F})$ is greater than the natural mortality $(\mathrm{M})$, and the exploitation rate $E>0.5$ is an indicator of overfishing.

## Karper/Barbs (Oseteochilus vittatus)

Based on length frequency data analysis of Karper/Barbs fish, the infinitive length $L_{\infty}=29 \mathrm{~cm}, K=0.55$ year $^{-1}, t_{0}=-0.30$, and the growth equation is $L_{t}=29\left(1-e^{-0.55(t+0.30)}\right)$. The number of samples ( $n$ ) taken for analysis was 2,593 fish (Figs. 6 and 7). Assuming an average temperature of $29.5^{\circ} \mathrm{C}$, based on Pauly's empirical equation, the natural mortality parameter value ( $M$ ) $=1.25$ year $^{-1}$. Based on the Jones and Van Zalinge plot analysis on the size composition of Karper/Barbs fish during the study,
the results obtained were total mortality $(Z)=3.7$ year $^{-1}$, fishing mortality $(F)=2.45$ year $^{-1}$ exploitation rate $(E)=0.66$. The fishing mortality $(F)$ is greater than the natural mortality $(M)$, and the exploitation rate $E>0.5$.

## Fish density based on depth stratification and target strength

The average density based on the depth stratification is $1-5$, $6-10,11-15$, and $>15 \mathrm{~m}$ (Table 3). The average value of fish density tended to be higher in the surface layer. The highest average density between $1-5 \mathrm{~m}$ beneath the surface was 20.7 Ind/ $\mathrm{m}^{3}$. In comparison, the lower fish density was identified in the deeper layer ( $>15 \mathrm{~m}$ ), reaching $0.2 \mathrm{Ind} / \mathrm{m}^{3}$, with an average of 5.4 Ind $/ \mathrm{m}^{3}$. Based on the acoustic survey results, small fish with a TS less than -47 dB were mainly detected in the surface layer, with a depth between $1-5 \mathrm{~m}$. This value gradually declined in the deeper layer. Meanwhile, fish with a higher TS value (TS > -47 dB ) tended to increase gradually according to depth and then decreased at a depth of more than 15 m (Table 4).

## Fish biomass based on reservoir depth and fish composition

The value of fish size ( cm ) should be converted to weight ( g ) by analyzing the relationship between length and weight of fish to determine the fish biomass. The fish selected for analysis of the relationship between length and weight was Nila/Cichlid (O. niloticus) because it is very dominant in Widas Reservoir and easy to catch from various sizes. Based on data analysis of length (cm) and weight $(\mathrm{g})$, the relationship between length and weight can be explained by $W=0.026 L^{2.904}$ (Fig. 8). After converting the fish size into weight, it can estimate fish biomass in the reser-


Fig. 4. The length frequency data of Nila/Cichlid (Oreochromis niloticus) in Widas Reservoir, East Java.


Fig. 5. Growth of Nila/Cichlid (Oreochromis niloticus) in Widas Reservoir, East Java.


Fig. 6. The length frequency data of Karper/Barbs (Osteochilus vittatus) in Widas Reservoir, East Java.


Fig. 7. Growth of Karper/Barbs (Osteochilus vittatus) in Widas Reservoir, East Java.

## Table 3. Average density of fish in each depth strata

| Depth | Ind $\left(\mathrm{m}^{3}\right)$ |
| :--- | :--- |
| $1-5 \mathrm{~m}$ | 20.7 |
| $6-10 \mathrm{~m}$ | 3.0 |
| $11-15 \mathrm{~m}$ | 0.6 |
| $>15 \mathrm{~m}$ | 0.2 |
| Average | 5.4 |

Table 4. Distribution of fish (\%) based on target strength and water depth

| Target strength $(\mathrm{dB})$ | Depth $(\mathrm{m})$ |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  | $1-5$ | $6-10$ | $11-15$ | $16-20$ |
| -55 | 4 | 8 | 0 | 0 |
| -54 | 22 | 4 | 0 | 0 |
| -53 | 37 | 17 | 0 | 0 |
| -52 | 19 | 17 | 0 | 0 |
| -51 | 11 | 17 | 0 | 0 |
| -50 | 0 | 13 | 0 | 0 |
| -49 | 7 | 8 | 9 | 0 |
| -48 | 0 | 0 | 0 | 0 |
| -47 | 0 | 4 | 0 | 0 |
| -46 | 0 | 4 | 0 | 0 |
| -45 | 0 | 8 | 21 | 15 |
| -44 | 0 | 0 | 14 | 43 |
| -43 | 0 | 0 | 14 | 14 |
| -42 | 0 | 0 | 21 | 14 |
| -41 | 0 | 0 | 14 | 14 |
| -40 | 0 | 0 | 0 | 0 |
| -39 | 0 | 0 | 7 | 0 |
| Total (\%) | 100 | 100 | 100 | 100 |

voir. Closer to the bottom of the reservoir, the channel becomes narrower. The surface area of the reservoir to a depth of 5 m was approximately 2.2 mile $^{2}, 6-10 \mathrm{~m}$ was 1.2 mile $^{2}, 11-15 \mathrm{~m}$ was about 0.4 mile $^{2}$, and the area of reservoir to the depth of $>15 \mathrm{~m}$ was 0.2 mile $^{2}$.

The estimated biomass was $58,813 \mathrm{~kg}$ for the depth of $1-5$ $\mathrm{m}, 11,998 \mathrm{~kg}$ for the depth of $6-10 \mathrm{~m}, 7,828 \mathrm{~kg}$ for the depth of $11-15 \mathrm{~m}$, and $1,219 \mathrm{~kg}$ for the depth of $15-20 \mathrm{~m}$, respectively. Therefore, the total biomass for the Widas Reservoir was approximately $79,848 \mathrm{~kg}$ or $165.67 \mathrm{~kg} / \mathrm{ha}$ (Table 5 and Fig. 9).

The catch of fish in Widas Reservoir using gillnet (mesh size $=2-3$ inch) was $5-8 \mathrm{~kg} /$ day/person, consisting of $78 \%$ Cichlid (Oreochromis), 13\% Minnows (Barbonymus), and 9\% Barbs (Osteochilus) (Fig. 10). The fish stock in the Widas Reservoir was about $79,848 \mathrm{~kg}$. If the composition of fish species in the stock follows the composition of the catch. Thus, the fish stock in the Widas Reservoir consists of Cichlid $62,281 \mathrm{~kg}$, Minnows $10,332 \mathrm{~kg}$, and Barbs $7,235 \mathrm{~kg}$.

The maximum sustainable yield (MSY) for the Cichlid was about $233,553 \mathrm{~kg}$, the Minnows group 24,280 kg, and the Barbs $13,384 \mathrm{~kg}$, respectively. The total allowable catch (TAC) for the Cichlid was approximately $186,842 \mathrm{~kg}$, Minnows $19,424 \mathrm{~kg}$, and Barbs $10,707 \mathrm{~kg}$. The exploitation rate $(E)$ for the Cichlid fish group was 0.83 , Minnows 0.74 , and Barbs 0.66 , respectively (Table 6).

## Discussion

Research on the study of fish stocks in inland waters, especially in East Java, still needs to be completed, making it difficult to document the potential of inland fisheries resources for their management. Information on growth, mortality, and exploita-


Fig. 8. Length-weight relationship of Nila/Cichlid (Oreochromis niloticus).


Fig. 9. Bathymetry and fish density in Widas Reservoir, East Java.

Table 5. Fish biomass in Widas Reservoir

| Depth | Biomass (kg) | Biomass $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$ | Biomass (kg/ha) |
| :--- | :--- | :--- | :--- |
| $1-5 \mathrm{~m}$ | 58,813 | $6,235.32$ | 62.35 |
| $6-10 \mathrm{~m}$ | 11,988 | $2,946.47$ | 29.46 |
| $11-15 \mathrm{~m}$ | 7,828 | $5,750.87$ | 57.51 |
| $>15 \mathrm{~m}$ | 1,219 | $1,634.46$ | 16.34 |
| Total | 79,848 | $16,567.11$ | 165.67 |

tion rate parameters is needed for managing fish resources (Raghavan et al., 2013). Effective inland fisheries management is constrained by the need for more introductory information on ecology, population, fish diversity, and endangered species (Cooke et al., 2012). Overfishing of freshwater fish has been the cause of declining fish stocks over the last few decades (Hicks \& McClanahan, 2012).

The results of this study indicate that fishing activities significantly impact the fish stocks in Widas Reservoir. Fishing


Fig. 10. Composition of fish catch in Widas Reservoir.

Table 6. Potency, total allowable catches of fish in widas reservoir

| Parameter of population | Total (kg/year) | Fish group |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  |  | Minnow | Cichlid | Barbs |
| MSY (kg/year) | 271,217 | 24,280 | 233,553 | 13,384 |
| TAC (kg/year) | 216,973 | 19,424 | 186,842 | 10,707 |
| Z (year ${ }^{-1}$ ) | 4.78 | 7.55 | 3.70 |  |
| F $\left(\right.$ year $\left.^{-1}\right)$ | 3.54 | 6.25 | 2.45 |  |
| M (year |  | 1.24 | 1.30 | 1.25 |
| E |  | 0.74 | 0.83 | 0.66 |

MSY, maximum sustainable yield; TAC, total allowable catch.
mortality is more significant than natural mortality in the three major fish in Widas Reservoir, indicating less influence of water pollution, such as MPs. According to Zhou et al. (2020), MP is pervasive and will have a long-term influence on the worldwide ecology. Nonetheless, the waters of Widas Reservoir are safe for fish growth and human health since they are still utilized as a drinking water source, meaning that the potency of water pollution is minimal in the study area (Aida et al., 2018). Moreover, there have been no reports or indicators that fish taken in Widas Reservoir have been contaminated with MPs. However, water quality issues should be further reconsidered in the fishery-based assessment. MPs is currently in the spotlight. It is being studied extensively in aquatic habitats worldwide, particularly their impact within a reservoir where the water motion is weak.

## Growth, mortality, and exploitation rate

Table 2 shows that fishing mortality $(F)$ of Tawes/Minnows ( $B$. gonionotus), Karper/Barbs (O. vittatus), and Nila/Cichlid ( $O$. niloticus) in the Widas Reservoir was more significant than the natural mortality $(M)$ and exploitation rate $(E)$ is already overfishing, $E>0.5$. These results are consistent with the assumption that a stock is optimally exploited when fishing mortality $(F)$ equals to natural mortality $(M)$, or $E=(F / Z)=0.5$. If $E>0.5$, it indicates that the fish resources have been overexploited.

The fishing mortality parameter of Tawes/Minnows was $(F)=3.54$ year $^{-1}$, and the exploitation rate $(E)=0.74$. Since $(E)$ $>0.5$, it shows the overfishing status. With the Von Bertalanffy equation, the infinity length $\left(L_{\infty}\right)$ of Tawes/Minnows fish ( $B$. gonionotus) was 45 cm , which will be reached at the age of more than seven years. Fish that are frequently caught at a size of $18-25 \mathrm{~cm}$ are thought to be 1-1.5 years old (Figs. 2 and 3). The environment strongly influences the growth of Tawes/Minnows fish (B. gonionotus). For instance, compared to the Pondok Reservoir, with a smaller area of 407 ha and worse water quality, the growth of Minnows is under the value estimated in the Widas Reservoir, whereby the fish growth is more negligible, with $\left(L_{\infty}\right)$ $=29 \mathrm{~cm}$ and $K=0.55$ year $^{-1}$ (Aida et al., 2022a).

According to King (2007), growth and death are influenced by ecological conditions. In general, the asymptotic difference in length $\left(L_{\infty}\right)$ and growth coefficient $(K)$ from one area to another can be caused by the availability of feed and water quality (Bartulović et al., 2004). Growth patterns are influenced by environmental factors, habitat, and food availability (Rahman et al., 2012). Several studies have shown that the parameters of the Tawes/Minnows fish population at each location differ due to different environmental conditions and the level of exploitation at each location. For example: Tawes/Minnows fish (Barbonymus schwanenfeldii) in Pedu Reservoir, Kedah Malaysia: ( $L_{\infty}$ ) $=30.95 \mathrm{~cm}$, growth coefficient $K=0.66$ year $^{-1}$, total mortality coefficient ( $Z$ ) is 2.01 year $^{-1}$, natural mortality coefficient $(M)$ is 1.37 year $^{-1}$, fishing mortality coefficient $(F)$ is 0.64 year $^{-1}$ and exploitation rate $(E)=0.32$ (Isa et al., 2012). Tawes/Minnows fish in Nagan River, Aceh Province, $L_{\infty}=16 \mathrm{~cm}$, growth coefficient $(K)=0.73$ year $^{-1}$, total mortality $(Z)=2.80$ year $^{-1}$, natural mortality $(M)=0.92$ year $^{-1},(F)=1.88$ year $^{-1}$, and exploitation rate $(E)$ $=0.67$ (Batubara et al., 2018).

Fishing activities for (B. gonionotus) in the Widas Reservoir cannot be more significant. The number of these fish needs to be a concern so that the population can stabilize and recover. The high level of exploitation of (B. gonionotus) in the Widas

Reservoir is due to the relatively large size of the fish, which makes it easy to catch and has substantial economic value as fish consumption (Utomo et al., 2019).

The infinity length of $O$. niloticus $\left(L_{\infty}\right)=44.4 \mathrm{~cm}$ is estimated to have reached the age of more than 6.5 years. O. niloticus, often caught between $15-25 \mathrm{~cm}$ in size, is estimated to be $0.5-1$ year old (Figs. 4 and 5). The exploitation rate $(E)=0.83$, meaning that the exploitation rate of $O$. niloticus in Widas Reservoir is already overfishing (Pauly, 1984). The high exploitation rate of $O$. niloticus in the Widas Reservoir is due to the relatively large size of this fish, catchable, and essential economic value as a consumption fish (Aida et al., 2022b). However, O. niloticus has a strong tolerance for environmental quality degradation and is easy to breed over time, so the population is more stable than B. gonionotus.

The growth of $O$. niloticus is strongly influenced by the environment; for example, the Rawa Pening Reservoir is more expansive, with the worse water quality than the Widas Reservoir, resulting in a slower growth, with an indication of a smaller infinity length $\left(L_{\infty}\right)=35 \mathrm{~cm}$ (Aida \& Utomo, 2016). O. niloticus in the Tamil Nadu Reservoir has an infinite length ( $L_{\infty}$ ) of 38.7 cm (male) and 37.6 cm (female), growth coefficient $(K)=0.65$ year ${ }^{-1}$ (male) and 0.73 year $^{-1}$ (female), fishing mortality $(\mathrm{F})=2.4$ year ${ }^{-1}$, the natural mortality $(\mathrm{M})=1.2$ year $^{-1}$ and the total mortality $(Z)=3.6$ year $^{-1}$, exploitation $(E)=0.67$. O. niloticus in the Sri Lanka Reservoir has an infinity length $\left(L_{\infty}\right)=42.8$ to 53.4 cm , growth constant $(K)$ between 0.20 to 0.49 year $^{-1}$, exploitation rate ( $E$ ) between ( 0.14 ) to ( 0.76 ) (Bandara et al., 2020).

Based on the Von Bertalanffy equation, the infinity length ( $L_{\infty}$ ) of Karper/Barbs fish (O. vittatus) was about 29 cm reaching the age of more than seven years. $O$. vittatus, often caught at a size of $10-15 \mathrm{~cm}$, are estimated to be $0.5-1$ year old. The maximum length ( $\mathrm{L}_{\text {max }}$ ) of O. vittatus caught was 27 cm , estimated to be more than five years old (Fig. 9). Exploitation rate was 0.66 (overfishing) (Pauly, 1984). Therefore, fishing activities for $O$. vittatus should not be excessive.

The existence of $O$. vittatus should be a concern so they can grow and breed properly. The high exploitation rate of $O$. vittatus in Widas Reservoir is due to its significant economic value, often caught by lift nets whose mesh size is less than 1.5 inches. The growth of $O$. vittatus is highly influenced by the environment. For example, in Rawa Pening Reservoir, although it is expansive, with the worse water quality compared to the Widas Reservoir, the growth of $O$. vittatus is lower with an indication of a small infinity length $\left(L_{\infty}\right)$ of 25 cm (Utomo et al., 2019).

According to RIFFE (2021), in general, fisheries exploitation in inland waters in Indonesia is already overfishing ( $\mathrm{E}>$ 0.5), except in Lake Limboto Sulawesi because the lake is vast and oligotroph, with no significant fishing activities (Table 7). According to Bandara et al. (2020), Charernnate et al. (2021), Feroz Khan et al. (2021), Johnson et al. (2020), Moreau \& Sricharoendham (1999); and Tesfaye et al. (2022), in general, fisheries exploitation in inland waters in several countries has been overexploited ( $\mathrm{E}>0.5$ ) (Table 8).

## Fish density and target strength

Larger fish are generally detected in deeper layers of water, while smaller fish are found in the surface layer (Utomo et al., 2019). The fish size significantly shapes the TS. Simmonds \& MacLennan (2005) stated that size is the dominant factor affecting the value of TS (TS). For the same fish species, the larger the fish size, the greater the TS value. The TS analysis may be more accurate if the relationship between TS and fish length is available. The most critical factor influencing acoustic estimates is the appropriate relationship between fish length and TS for a given fish population (Frouzova et al., 2005).

Simmonds \& MacLennan (2005) stated that several other factors, including swimming bubbles, fish behavior, acoustic impedance, sound wavelength, beam pattern, fish species, and swimming speed, influence the TS value. Three other factors contributing to the vertical distribution of the target force were temperature $19.87 \%$, salinity $12.61 \%$, and depth $61 \%$. In addition, disturbances from ships, such as ship movement and noise from generators and propellers, cause fish to avoid vessels (vessel avoidance) to the deeper water column (Fauziyah, 2010).

The behavior of the fish influences TS, the angle of the fish swimming to the hydroacoustic axis, and the angle is the angle of inclination or incidence. At night, when the fish creeps, the inclination angle approaches zero (Brooking \& Rudstam, 2011). TS estimation experiments in cages have certain advantages in developing TS, as they are easy to carry out. For freshwater fish, it is easier to do TS experiments and freshwater fish can be adjusted so that their body position does not change, so it will be easier to research TS on fish (Mouget et al., 2019). Fish that do not move, for example, killed or stunned, are easier to test for TS.

Getting to the bottom of the water shows the increasing number of fish with greater TS. In general, larger fish were detected at deeper depths. This state relates to the differences in swimming layers for each size of fish. Larger fish tend to swim

Table 7. Comparison of population parameters of Bonionotus gonionotus, Osteochilus vittatus, and Oreochromis niloticus in several areas throughout Indonesia

| Location | Fish species | $L_{\infty}(\mathrm{cm})$ | $K\left(y^{\text {ear }}{ }^{-1}\right)$ | $M\left(\right.$ year $\left.^{-1}\right)$ | $Z\left(\right.$ year $\left.^{-1}\right)$ | $F\left(\right.$ year ${ }^{-1}$ ) | E | References |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Papua Island, Ayamaru Lake | B. gonionotus | 40.50 | 0.30 | 0.55 | 1.65 | 1.10 | 0.66 | RIFFE (2021) |
|  | O. niloticus | 26.50 | 0.30 | 0.75 | 2.99 | 2.24 | 0.75 |  |
|  | O. vittatus | 24.00 | 0.20 | 0.74 | 1.11 | 0.37 | 0.34 |  |
| Sulawesi Island, Limboto Lake | B. gonionotus | 32.00 | 0.36 | 0.90 | 1.32 | 0.42 | 0.32 | RIFFE (2021) |
|  | O. niloticus | 25.90 | 0.29 | 0.82 | 1.14 | 0.31 | 0.28 |  |
| Maluku Island, Batu Jai Reservoir | B. gonionotus | 27.00 | 0.41 | 1.06 | 2.19 | 1.13 | 0.52 |  |
|  | O. niloticus | 31.00 | 0.55 | 1.23 | 2.66 | 1.43 | 0.54 |  |
| Kalimantan Island, Baritp River | B. gonionotus | NA | NA | 0.75 | 1.95 | 1.20 | 0.62 | RIFFE (2021) |
|  | O. niloticus | NA | NA | 1.24 | 2.87 | 1.63 | 0.57 |  |
|  | O. vittatus | NA | NA | 2.51 | 3.4 | 0.88 | 0.26 |  |
| Sumatera Island, Ranau Lake | B. gonionotus | 47.00 | 0.21 | 0.93 | 2.92 | 1.98 | 0.68 | RIFFE (2021) |
|  | O. niloticus | 36.00 | 0.17 | 0.51 | 0.84 | 0.33 | 0.40 |  |
|  | O. vittatus | 25.00 | 0.44 | 1.12 | 2.06 | 0.94 | 0.46 |  |
| Sumatera Island, Aceh Nagan River | B. gonionotus | 16.00 | 0.73 | 0.92 | 2.80 | 1.88 | 0.67 | Batubara et al. (2018) |

NA, not applicable.

Table 8. Comparison of population parameters of Bonionotus gonionotus, Osteochilus vittatus, and Oreochromis niloticus based on studies in other countries

| Species | Location | $L_{\infty}(\mathrm{cm})$ | $K\left(Y_{\text {ear }}{ }^{-1}\right)$ | $t_{0}$ (Year) | $M\left(\mathrm{Year}^{-1}\right)$ | $Z\left(\right.$ Year $\left.^{-1}\right)$ | $F\left(\right.$ Year ${ }^{-1}$ ) | E | References |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| O. niloticus | Rajanganaya, Srilangka | 53.4 | 0.49 | -0.29 | 0.94 | 3.84 | 2.90 | 0.76 | Bandara et al. (2020) |
|  | Halali Reservoir, India | 46.73 | 0.63 | - | 0.60 | 1.5 | 0.72 | 0.48 | Johnson et al. (2020) |
|  | Kelavarappalli Reservoir, Tamil Nadu | 38.7 | 0.65 | - | 1.2 | 3.6 | 2.4 | 0.67 | Feroz Khan et al. (2021) |
|  | Lake Langeno, Ethiopia | 35.70 | 0.32 | -0.49 | 0.82 | 2.31 | 1.56 | 0.67 | Tesfaye et al. (2022) |
| Barbonymus sp. (Minnows) | Kangkrajan Reservoir, Thailand | 36.2 | 0.39 | -0.28 | 0.93 | 2.44 | 1.51 | 0.55 | Charernnate et al. (2021) |
| O. vittatus | Rajjaprabha Reservoir, Thailand | 34.5 | 0.32 | - | 0.81 | 2,25 | 1.44 | 0.64 | Moreau \& Sricharoendham (1999) |

at a deeper depth than smaller fish. Some of the advantages in freshwater systems are that the sound emission from the transducer is more comprehensive, indicating lower susceptibility to noise and increased resistance to fish aspects (Kang \& Hwang, 2003).

## Fish biomass in Widas Resevoir

Based on the results of the bathymetry mapping carried out in June 2019, the depth of the Widas Reservoir ranged from 2 to 20 m . The deepest part of the Widas Reservoir ( 20 m ) is situated in the middle of the reservoir and near the outlet. Meanwhile, a depth of $0.5-2 \mathrm{~m}$ is often found within the inlets. Fish
biomass in the surface layer ( $1-5 \mathrm{~m}$ ) is more significant than in the deeper layer by more than 15 m (Table 5) because the area and volume of water towards the bottom are getting narrower. The oxygen content is also lowered with the increase in the water depth. Oxygen is essential for fish and other biota, so fish biomass declines towards the bottom (Utomo et al., 2019). The shallow waters are generally located on the reservoir's edge, with sufficient natural food for fish, especially aquatic vegetation, insects, periphyton, and others. Therefore, fish biomass is more abundant (Aida et al., 2022b). Freshwater fish generally spawn on the banks of rivers, reservoirs, and lakes during the rainy season. Aquatic plants such as grass on the shores of lakes/res-
ervoirs are good places as nursery grounds for young fish (Djumanto et al., 2020).

Getting to the bottom of the waters, it is inhabited by fish detected with a higher dB (decibel) value (Table 4), and this is an indication that the deep waters are inhabited by large fish (Fauziyah, 2010). Many fish are in the surface layer because small fish mostly prey for plankton such as Wader (Rasbora), Karper (Osteochilus), and others (Ma3mun et al., 2022). Meanwhile, large fish such as Patin (Pangasius), Tapa (Wallago), Toman (Channa), and others require deep waters, especially during the dry season, such as the trough of the river (Utomo et al., 2019). Small fish, in general, often live in shallow reservoir borders because there is much aquatic vegetation, such as shelters and bays with calm waters, which are suitable habitats for fish. Pelagic fish migrate towards the surface at night and inhabit the surface column at night. Pelagic fish migrate to adapt to environmental factors such as temperature, salinity, and oxygen levels (Kaartvedt et al., 2007). The vertical distribution of fish in the lake is mainly affected by water temperature and oxygen stratification, but other factors, such as the risk of evading predation (Tsagarakis et al., 2015).

All exploitation rates $(E)$ are more significant than 0.5 , indicating overfishing (Table 6). Based on production data, fish catch in the Widas Reservoir reached $283,000 \mathrm{~kg} / \mathrm{year}$, while the TAC was only $216,973 \mathrm{~kg} /$ year, proving the overfishing state.

On the other hand, Cichlid fish dominate the Widas reservoir. These fish are easy to breed in various types of waters and even tend to be invasive. Native fish often cannot compete in breeding, competing for food and habitat (Syafei \& Sudinno, 2018). Cichlid fish have significantly developed in Indonesian waters, especially reservoirs and lakes. Some research results show that Cichlid fish have dominated the reservoir waters. For example, in Gajah Mungkur Reservoir, the production of Cichlid catches reached $40 \%$. In Sragen Regency, it reached $58.9 \%$; in Cikoncang Reservoir, it reached $58.5 \%$. Cirata Reservoir reached $53 \%$; Pondok Reservoir reached 76\% (Aida et al., 2018).

After obtaining the value of fish stocks, population parameters, and TAC, it is expected to be a reference material for fish resource management policies in Widas Reservoir. There are already indications of overfishing, so fishing activities must be reduced, especially gillnet and lift net, which are more numerous. Lift-net is a non-selective fishing gear in East Java and Central Java, and the number continues to increase in almost all reservoirs. Many gillnets used by fishers are also not selective in the
size of their nets under 2 cm , so the fish that have not spawned have been caught significantly. There are many fishing activities at the reservoir inlet at the beginning of the rainy season and fish spawning ground at the beginning of the rainy season. It is necessary to introduce closed seasons or protected areas at the inlet of reservoirs (Aida et al., 2018).

In addition to regulating fishing and conservation, fish stocking must be carried out continuously. Restocking fish from hatcheries (culture-based fisheries, or CBF) is one approach to increase capture fisheries yields; CBF should be implemented in Indonesia using local species (Utomo et al., 2019). A Reservoir is a good body of water for restocking because reservoirs have clear boundaries that make them easier to manage (Aida et al., 2022b). The stocking of fish should be convinced that there is a compatibility between the fish and its aquatic environment, a spawning habitat, enough natural food available, and fishing rules; if necessary, a conservation area is created (Utomo, 2016). Future research needed in Widas Reservoir is to support restocking, conservation, and fishing rules so that fish populations in Widas Reservoir can increase (Utomo \& Aida, 2017).

## Conclusion

The fish stock in Widas Reservoir is $79,848 \mathrm{~kg}$ and is dominated by Cichlid species. The stock gradually declines with the increase in water depth. The fish stock in the surface layer at a depth between $1-5 \mathrm{~m}$ is $58,813 \mathrm{~kg}$, while in the deeper layer (> 15 m ), the stock only reaches $1,219 \mathrm{~kg}$. Based on the analysis of the dominant fish growth parameter data in the Widas Reservoir, the indication of overfishing is possible since the exploitation rate value was generally more than 0.6 . The overfishing state is proven by the significantly higher catch production than TAC, with a deviation of $66,017 \mathrm{~kg} /$ year in the Widas Reservoir.

It is necessary to regulate fishing in the Widas Reservoir so that the catch does not exceed the TAC. Fishing activities must be more controlled, mainly controlling the usage of gillnet and lift nets with mesh sizes of less than 2 inches, which have been increasing recently. In addition to the fish population, it is necessary to restock the native fish, such as B. gonionotus, and $O$. vittatus, for sustainability. A further assessment of fish restocking, conservation, and decision-making is necessary to recover the fish population and control the overfishing issue.

## Competing interests

No potential conflict of interest relevant to this article was re-
ported

## Funding sources

Not applicable.

## Acknowledgements

The authors would like to thank Research Institute for Inland Fisheries and Fisheries Extension Palembang for all supports during this research implementation, research team colleagues who have helped during the research and the fishers as enumerators who helped in the data collection.

## Availability of data and materials

Upon reasonable request, the datasets of this study can be available from the corresponding author.

## Ethics approval and consent to participate

This article does not require IRB/IACUC approval because there are no human and animal participants.

## ORCID

Siti Nurul Aida https://orcid.org/0000-0003-1595-6331

Agus Djoko Utomo
Safran Makmur
Tuah Nanda M. Wulanda

Khoirul Fatah
Yosmaniar
Indra Suharman
Ulung Jantama Wisha
https://orcid.org/0000-0003-4185-5980
https://orcid.org/0000-0002-4256-909X
https://orcid.org/0000-0002-2459-956X
https://orcid.org/0000-0001-7428-5385
https://orcid.org/0000-0001-9931-2669
https://orcid.org/0000-0003-4314-2458
https://orcid.org/0000-0001-6064-675X

## References

Aida SN, Ridho MR, Saleh E, Utomo AD. Population dynamics of parameters and the size of the first gonad maturity on Java barbs (Barbsonymus gonionotus) in Pondok reservoir, East Java. Ecol Environ Conserv Pap. 2022a;28:89-94.
Aida SN, Utomo AD. Assessment of water quality for fisheries in Rawa Pening Centre of Java. Bawal. 2016;8:173-82.
Aida SN, Utomo AD. Carrying capacity estimation for fish culture of floating net cages in Pondok reservoir, Ngawi East Java. Bawal. 2018;10:197-208.
Aida SN, Utomo AD, Adjie S, Subroto G, Waro B. Technical reports of fish stock and fishery potential, East Java (KPPPUD 431). Palembang: Research Institute for Inland Fish-
eries and Fisheries Extensions; 2018. p. 133.
Aida SN, Utomo AD, Anggraeni DP, Ditya YC, Wulandari TNM, Ali M, et al. Distribution of fish species in relation to water quality conditions in Bengawan Solo river, Central Java, Indonesia. Pol J Environ Stud. 2022b;31:5549-61.
Bandara KVSN, Ajith Kumara PAD, Amarasinghe US. Population dynamics of Nile tilapia, Oreochromis niloticus (Linnaeus, 1758) (Teleostei, Cichlidae), in some irrigation reservoirs of Sri Lanka. Asian Fish Sci. 2020;33:192-204.
Bartulović V, Glamuzina B, Conides A, Dulčić J, Lučić D, Njire J, Kožul V. Age, growth, mortality and sex ratio of sand smelt, Atherina boyeri Risso, 1810 (Pisces: Atherinidae) in the estuary of the Mala Neretva River (middle-eastern Adriatic, Croatia). J Appl Ichthyol. 2004;20:427-30.
Batubara AS, Muchlisin ZA, Efizon D, Elvyra R, Fadli N, Irham M. Morphometric variations of the genus Barbonymus (Pisces, Cyprinidae) harvested from Aceh Waters, Indonesia. Fish Aquat Life. 2018;26:231-7.
Brooking TE, Rudstam LG. Hydroacoustic target strength distributions of alewives in a net-cage compared with field surveys: deciphering target strength distributions and effect on density estimates. Trans Am Fish Soc. 2011;138:471-86.
Buwono NR, Risjani Y, Soegianto A. Distribution of microplastic in relation to water quality parameters in the Brantas river, East Java, Indonesia. Environ Technol Innov. 2021;24:101915.
Charernnate K, Noranarttragoon P, Jutagate T. Length-based stock assessment of Smith's Barb, Puntioplites proctozystron (Bleeker, 1865) (Cyprinidae) and Asian redtail catfish, Hemibagrus nemurus, (Valenciennes, 1840), (Bagridae) in a multipurpose reservoir in Thailand. Asian Fish Sci. 2021;34:159-67.
Cooke SJ, Paukert C, Hogan Z. Endangered river fish: factors hindering conservation and restoration. Endanger Species Res. 2012;17:179-91.
Djumanto, Setyobudi E, Simanjuntak CPH, Fadjar Rahardjo M. Estimating the spawning and growth of striped snakehead Channa striata Bloch, 1793 in Lake Rawa Pening Indonesia. Sci Rep. 2020;10:19830.
Fauziyah JA. Acoustic density of small pelagic fish in the Arafura sea. J Sci Res. 2010;13:13106.
Feroz Khan M, Panikkar P, Das BK. Population dynamics of Nile tilapia Oreochromis niloticus (Linnaeus, 1758) at Kelavarappalli reservoir in Tamil Nadu. J Inland Fish Soc India. 2021;53:57-65.

Frouzova J, Kubecka J, Balk H, Frouz J. Target strength of some European fish species and its dependence on fish body parameters. Fish Res. 2005;75:86-96.
Gayanilo FC Jr, Sparre P, Pauly D. The FAO-ICLARM stock assessment tools (FISAT) user's guide. In: FAO computerized information series fisheries. Rome: FAO; 1995. p. 126.
Hannachi MS, Ben Abdallah L, Marrakchi O. Acoustic identification of small-pelagic fish species: target strength analysis and school descriptor classification. MedSudMed Tech Doc. 2004;5:90-9.
Hart BT, van Dok W, Djuangsih N. Nutrient budget for Saguling reservoir, West Java, Indonesia. Water Res. 2002;36:215260.

Hicks CC, McClanahan T. Assessing gear modifications needed to optimize yields in a heavily exploited, multi-species, seagrass and coral reef fishery. PLOS ONE. 2012;7:e36022.
Isa MM, Md-Shah AS, Mohd-Sah SA, Baharudin N, AbdulHalim MA. Population dynamics of tinfoil barb, Barbonymus schwanenfeldii (Bleeker, 1853) in Pedu Reservoir, Kedah. J Biol Agric Healthc. 2012;2:55-70.
Johnson C, Sarkar UK, Koushlesh SK, Das AK, Das BK, Naskar BK. Population structure of Nile tilapia and its impact on fisheries of a tropical impacted reservoir, Central India. Environ Sci Pollut Res. 2020;27:29091-9.
Kaartvedt S, Klevjer TA, Torgersen T, Sørnes TA, Røstad A. Diel vertical migration of individual jellyfish (Periphylla periphylla). Limnol Oceanogr. 2007;52:975-83.
Kang D, Hwang D. Ex situ target strength of rockfish (Sebastes schlegeli) and red sea bream (Pagrus major) in the Northwest Pacific. ICES J Mar Sci. 2003;60:538-43.
King M. Fisheries biology, assessment and management. Oxford: Blackwell; 2007. p. 382.
Lian Y, Huang G, Godlewska M, Cai X, Li C, Ye S, et al. Hydroacoustic estimates of fish biomass and spatial distributions in shallow lakes. J Oceanol Limnol. 2017;36:587-97.
Mảmun A, Nurdin E, Priatna A, Mahiswara M. Relationship between the existence of tuna to upper and lower thermocline at fads in Pelabuhan Ratu, West Java. Indones Fish Res J. 2022;28:1-13.

Mello LGS, Rose GA. The acoustic dead zone: theoretical vs. empirical estimates, and its effect on density measurements of semi-demersal fish. ICES J Mar Sci. 2009;66:1364-9.
Moreau J, Sricharoendham B. Growth, mortality and recruitment of fish populations in an Asian man made lake Rajjaprabha reservoir (Thailand) as assessed by length fre-
quency analysis. Asian Fish Sci. 1999;12:277-88.
Mouget A, Goulon C, Axenrot T, Balk H, Lebourges-Dhaussy A, Godlewska M, et al. Including 38 kHz in the standardization protocol for hydroacoustic fish surveys in temperate lakes. Remote Sens Ecol Conserv. 2019;5:332-45.
Nejad JG, Ataallahi M, Park KH. Methodological validation of measuring Hanwoo hair cortisol concentration using bead beater and surgical scissors. J Anim Sci Technol. 2019;61:41-6.
Pauly D. Some simple methods for the assessment of tropical fish stocks. Rome: FAO Fisheries \& Aquaculture; 1984. p. 52.

Perivolioti TM, Frouzova J, Tušer M, Bobori D. Assessing the fish stock status in lake Trichonis: a hydroacoustic approach. Water. 2020;12:1823.
Prabowo AD, Nawiyanto S. Pembangunan irigasi widas dan pengaruhnya terhadap ekonomi dan lingkungan di nganjuk, Jawa Timur 1978-2010. Handep J Sejarah Budaya. 2020;4:19-38.
Raghavan R, Dahanukar N, Tlusty MF, Rhyne AL, Krishna Kumar K, Molur S, et al. Uncovering an obscure trade: threatened freshwater fishes and the aquarium pet markets. Biol Conserv. 2013;164:158-69.
Rahman MA, Arshad A, Amin, SMN. Growth and production performance of threatened snakehead fish, Channa striatus (Bloch), at different stocking densities in earthen ponds. Aquac Res. 2012;43:297-302.
Renfree JS, Demer DA. Optimizing transmit interval and logging range while avoiding aliased seabed echoes. ICES J Mar Sci. 2016;73:1955-64.
RIFFE. Technical reports of fish stock and fishery potential, East Java (KPP-PUD). Palembang: Research Institute for Inland Fisheries and Fisheries Extension; 2021. p. 133.
Simmonds J, MacLennan D. Fisheries acoustics theory \& practice. In: Fish \& aquatic resources series 10. 2nd ed. London: Blackwell Publication; 2005.
Sparre P, Venema SC. Introduction to tropical fish stock assessment: part 1. manual. Rome: FAO Fisheries \& Aquaculture; 1992. p. 376.

Susilo H. Management optimization on pondok reservoir in Ngawi, East Java. Pilar Teknol. 2019;10:54-8.
Syafei L, Sudinno D. Invasive alien spesies, sustainability aquatic biodiversity challenges. J Fish Mar Ext. 2018;12:145-61.
Tesfaye G, Tesfaye G, Getahun A, Tadesse Z, Workiye G. Population dynamics of the Nile tilapia (Oreochromis niloticus L.
1758) stock in Lake Langeno, Ethiopia. SINET Ethiop J Sci. 2022;45:174-91.
Tsagarakis K, Giannoulaki M, Pyrounaki MM, Machias A. Species identification of small pelagic fish schools by means of hydroacoustics in the Eastern Mediterranean Sea. Mediterr Mar Sci. 2015;16:151-61.
Utomo AD. Optimization of fisheries reserve management in floodplain for the preservation of fish resources. Research professor inauguration speech. Jakarta: Indonesian Institute of Sciences (LIPI); 2016.
Utomo AD, Samuel. Sustainable management of inland capture fisheries for food security: experience of Indonesia. Fish People. 2017;15:28-35.
Utomo AD, Wibowo A, Suhaimi RA, Atminarso D, Baumgartner LJ. Challenges balancing fisheries resource management and river development in Indonesia. Mar Freshw Res. 2019;70:1265-73.
Viani DZ, Retnaningdyah C. Evaluation of trophic status and organic pollution at Lahor reservoir Malang using diatoms as bioindicator. Biotropika. 2018;6:10-5.
Wheeland LJ, Rose GA. Quantifying fish avoidance of small acoustic survey vessels in boreal lakes and reservoirs. Ecol Freshw Fish. 2014;24:67-76.
Zhou A, Zhang Y, Xie S, Chen Y, Li X, Wang J, et al. Microplastics and their potential effects on the aquaculture systems: a critical review. Rev Aquac. 2020;13:719-33.


[^0]:    Received: Oct 28, 2023 Revised: Dec 13, 2023 Accepted: Dec 28, 2023
    ${ }^{*}$ Corresponding author: Ulung Jantama Wisha
    Research Center for Oceanography, National Research and Innovation Agency (BRIN), Jakarta 14430, Indonesia
    Tel: +62-2164713850, E-mail: ulun002@brin.go.id
    This is an Open Access article distributed under the terms of the Creative Commons Attribution Non-Commercial License (http://creativecommons.org/licenses/by$\mathrm{nc} / 4.0 /$ ) which permits unrestricted non-commercial use, distribution, and reproduction in any medium, provided the original work is properly cited. Copyright © 2024 The Korean Society of Fisheries and Aquatic Science

[^1]:    GT, gross tone.

